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Species-specific responses of marine bacteria to environmental perturbation

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Environmental perturbations shape the structure and function of microbial communities. Oil spills are a major perturbation and resolving spills often requires active measures like dispersant application that can exacerbate the initial disturbance. Species-specific responses of microorganisms to oil and dispersant exposure during such perturbations remain largely unknown. We merged metatranscriptomic libraries with pangenomes to generate **Core-Accessory Metatranscriptomes (CA-Metatranscriptomes)** for two microbial hydrocarbon degraders that played important roles in the aftermath of the Deepwater Horizon oil spill. The *Colwellia* CA-Metatranscriptome illustrated pronounced dispersant-driven acceleration of core (~41%) and accessory gene (~59%) transcription, suggesting an opportunistic strategy. *Marinobacter* responded to oil exposure by expressing mainly accessory genes (~93%), suggesting an effective hydrocarbon-degrading lifestyle. The CA-Metatranscriptome approach offers a robust way to identify the underlying mechanisms of key microbial functions and highlights differences of specialist-vs-opportunistic responses to environmental disturbance.

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INTRODUCTION

The ocean is the Earth's oldest and most dynamic habitat. Microbes form the base of the ocean's food web and moderate key ecosystem services [1, 2], producing half of the Earth's oxygen each year, sequestering carbon via the primary fixation of inorganic carbon into biomass, modulating global biogeochemical cycles, and transforming and detoxifying pollutants [3]. Marine microbial populations are taxonomically and functionally diverse [4]. The low abundant "rare biosphere" accounts for a majority of the phylogenetic diversity in microbial populations [5, 6] and represents a tremendous repository of metabolic functionality [7]. Across diverse systems such as the human microbiome [8], soils [9, 10], the gut microbiome of soil fauna [11], aquatic systems [12, 13], and marine sediments [14], disturbances influence microbial community structure and function in ways that impact the ecosystem services that microorganisms provide [15]. Identifying why certain taxa respond to disturbances more effectively than others represents a key step towards achieving predictability of the functional capacity of microbial populations and the ecosystem dynamics they mediate. An improved understanding of the mechanisms by which key members of microbial communities respond to disturbance could provide a unifying framework, bridging biogeochemistry, microbial ecology, and related scientific disciplines [16].

Oil spills are a frequent perturbation to marine – and other – environments, yet much remains to be learned about how these perturbations alter oceanic microbial communities [17]. We explored data generated during a laboratory simulation of an oil spill [18] to identify how key members of the microbial rare biosphere responded to different environmental conditions. In April 2010, a catastrophic explosion on the Deepwater Horizon drilling rig sank the platform, initiating an uncontrolled discharge of 4.9 million barrels of crude oil into the Gulf of Mexico over 84 days. A key active measure to the oil spill was the application of seven million liters of synthetic dispersants (Corexit EC9500A and EC9527A) to surface oil slicks, and to the discharging wellhead at 1500 m water depth [19, 20]. Dispersant application enhanced oil droplet formation and aqueous phase solubilization of oil with a secondary aim of stimulating biodegradation [18]. However, dispersants had negative effects on microbial communities [18, 21, 22]. A more comprehensive understanding of how dispersants impact microorganisms involved in hydrocarbon oxidation may reveal how key microorganisms respond to disturbance, providing key information necessary to predict the trajectory and rate of recovery of baseline microbial communities in the aftermath of major disturbances. The approach presented here – merging metatranscriptomics with pangenomic data – is a

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robust first step to identify how individual species respond to perturbation and promotes a mechanistic understanding of impact(s) of environmental disturbance. We assessed the response of key oil-degrading taxa to dispersant and/or oil exposure in microcosms and found that patterns of ecological succession in the microcosms mirrored those observed in the deepwater oil plumes during the Deepwater Horizon oil well blowout [23–25]. Indigenous hydrocarbon degraders occupy the rare biosphere of Gulf waters [7], and are primed to respond to hydrocarbon inputs from natural or anthropogenic sources [17]. Seven weeks after the Deepwater Horizon incident began, the community dominated by *Oceanospirillales* shifted to a community characterized by abundant *Cycloclasticus* and *Colwellia* [26–28]. *Colwellia* is a genus of psychrophilic marine bacteria [29], first isolated from the Puerto Rico Trench. *Colwellia* has not been shown to degrade oil or oil components, but a single amplified genome possessed the genes necessary to degrade alkanes and aromatic hydrocarbons [30]; *Colwellia* can also produce unique and relevant glycoproteins and natural surfactants [31, 32]. *Colwellia* spp. responded rapidly in situ during the Deepwater Horizon incident [28, 33], and in experiments where Gulf seawater was amended with oil, dispersed oil, or dispersant [18, 28, 33, 34]. *Marinobacter*, first described in 1992 [35], colonizes a broad range of marine habitats [36–41] and utilizes a remarkable range of electron acceptors and electron donors [42]. *Marinobacter* strains degraded alkanes [43] and polycyclic aromatic hydrocarbons (PAH) under anoxic conditions during the Deepwater Horizon incident [44]. Chemical dispersants inhibit *Marinobacter* sp. [18, 21, 22, 45]. Recently, Rughöft et al. showed reduced growth and hydrocarbon biodegradation in starved cultures of *Marinobacter* sp. TT1 following Corexit EC9500A exposure, suggesting that substrate availability and nutritional status poises its response to environmental perturbation [37]. The ecological fitness of microorganisms such as *Marinobacter* and *Colwellia* may influence their response to environmental perturbation, especially to oil and dispersant exposure.

We examined the transcriptional signatures from lab microcosm experiments [18] to investigate the physiological drivers of niche partitioning in *Colwellia* and *Marinobacter* through assessing Differentially Expressed (DE) genes in the context of their pangenomes. We refer to the mapping of metatranscriptomic signals in the context of core versus accessory genome as the “CA-Metatranscriptome”. Application of this approach revealed distinct modes of response to environmental perturbation and offers a robust and powerful way to track and understand species-specific responses to acute and chronic perturbations in other systems.

MATERIALS AND METHODS

Initial experiments

The procedures and protocols for the experiments were presented previously [18, 46, 47]. Briefly, Kleindienst et al. [18] simulated environmental conditions in the oil-rich 1100 m deep water plume that formed during the Deepwater Horizon spill. The experiment tracked the impacts of oil-only (supplied as a water-accommodated fraction, hereafter “WAF”), synthetic dispersant (Corexit 9500; “DISP”), oil-Corexit mixtures (chemically enhanced water-accommodated fraction, “CEWAF”) and chemically enhanced water-accommodated oil fraction with nutrients (hereafter, “CEWAFN”) additions to naturally-occurring microbial populations in deep-water collected from a natural seep site in the Gulf of Mexico (Green Canyon lease block 600, e.g., GC600). The experiment was designed to reveal the impact of distinct exposure regimes resulting from infusions of organic carbon derived from oil, synthetic dispersant, or oil and synthetic dispersant on microbial community evolution and activity over time. Exposure to different organic carbon regimes is known to drive diverging patterns of microbial community composition and activity [48]. Kleindienst et al. [18] observed that exposure to synthetic dispersant alone did not enhance heterotrophic microbial activity or hydrocarbon oxidation rates [18] but it did increase the abundance of *Colwellia*, key players during the Deepwater Horizon oil spill as evidenced by their enrichment in the deep-sea plume [7]. In contrast, exposure to oil, but not synthetic dispersant, increased the abundance of

Marinobacter. In addition, hydrocarbon oxidation rates were elevated in the presence of WAF. Here, we expand on the findings of Kleindienst et al. [18] by analyzing and interpreting metatranscriptomic data from their experiment and by applying the CA-Metatranscriptome approach to shed light on the mechanisms employed by the key microbial players *Colwellia* and *Marinobacter* to respond to oil and dispersant exposure. Further analyses from the broader metatranscriptomic assessment is presented by Peña-Montenegro et al (submitted) [47].

Data workflow and background pangenomes

This work involved the following steps: A) the curation of a pangenome for each of the target organisms (i.e., *Colwellia* and *Marinobacter*); B) the processing of transcriptomic signals and statistical tests to determine DE genes; and C) the generation of CA-Metatranscriptomes.

We followed a previously reported protocol for the pangenomic analysis [49]. All available complete and near-to-complete genomes in NCBI of *Colwellia* ($n = 77$) and *Marinobacter* ($n = 171$) (Supplementary Tables 1 and 2) were processed in Anvi'o (v.7.1) [50]. Anvi'o genome databases were generated with the *anvi-gen-genomes-storage* program to store all genomes and associated annotations. Each pangenome was computed with the *anvi-pan-genome* program to identify gene clusters. Background pangenomes were later used to generate CA-Metatranscriptomes (see below).

RNA sequencing

RNAseq library preparation, sequencing, and further data processing details are described in Peña-Montenegro et al (submitted) [47] and all scripts are found on Github (<https://github.com/biotemon/K2015>). Anvi'o can accept only one reference genome and one or more (meta)transcriptomes to calculate appropriate read recruitment layers for each gene cluster. To fulfill the “one reference genome” requirement, we identified the best genome reference using an all-against-all recruitment screening approach, as published previously [51–53]. All metatranscriptomic libraries were aligned to all available complete and near-to-complete genomes in NCBI of *Colwellia* ($n = 77$) and *Marinobacter* ($n = 171$) (Supplementary Tables 1 and 2). The largest mapping recruitment (i.e., average mapping counts and average mapping rates) was used to identify the best reference genome from NCBI for each species. For *Colwellia*, ~13.4 million reads were recruited for the Metagenome Assembled Genome (MAG) assigned to Candidatus *Colwellia aromaticivorans*, which we refer as *Colwellia* MAG-2 (Accession Number: QOLD00000000.1) henceforth. For *Marinobacter*, ~1.7 million reads were recruited by *Marinobacter* sp. C18 (Accession Number: LQXJ00000000.1) (Supplementary Data 1). SAM files were sorted and indexed into BAM files for later processing in the DE analysis and CA-Metatranscriptome generation.

Differential expression analysis was processed in R (v.4.3.0) as follows. Transcript counts were extracted by the mapping files using *ht-seq* and subsequently normalized using the Transcripts per Million (TPM) approach, defined as the relative abundance of an expressed gene maintaining comparability across samples and treatments [54]. TPM profiles were processed using *DESeq2* (v.1.26.0) using the regularized logarithm transformation [55]. Contrasts with respect to the biotic control (i.e., treatments without amendments) were performed using the negative binomial Wald test with Cook's distance to control for outliers. Genes with an adjusted p -value < 0.05 determined using the Benjamini–Hochberg method were classified as DE genes. Mapping counts and normalized scores for all DE genes are shown in Supplementary Data 2.

Generating CA-metatranscriptomes

To explore ecological implications of DE genes in the context of the *Colwellia* and *Marinobacter* pangenomes, we followed the workflow outlined in reference [49] and <http://merenlab.org/2016/11/08/pangenomics-v2/>, with minor modifications. Additional layers including the detection of DE genes in gene clusters were included to the pangenomic databases using the program *anvi-import-misc-data*. We visualized the CA-Metatranscriptome using the program *anvi-display-pan*. Using the *anvi-display-pan* interactive interface we searched for all gene clusters detected in all the genomes to assign them into the bin CORE. All scripts are found on Github (<https://github.com/biotemon/K2015>). Anvi'o procedures for *Colwellia* and *Marinobacter* CA-Metatranscriptomes are described in Supplementary Data 3 and 4, respectively.

Pathway visualization for Fig. 3 was generated using the iPath 3 module based on KEGG orthology numbers. An interactive version of the metabolic map is available at <https://pathways.embl.de/ipath3.cgi?s=xqvYg5kHHCSxGqWy8yJ>.

RESULTS

Kleindienst et al. [18] demonstrated differences in the response of *Colwellia* and *Marinobacter* abundance to specific cocktails of organic carbon [46] (Materials and Methods). Four triplicated microcosms (DISP, WAF, CEWAF, CEWAFN) and triplicated biotic controls were sampled over a span of six weeks (at 0, 7, 17, 28, and 42 days, except CEWAFN treatments, which were only sampled at 0, 7, and 42 days) and transcriptomic libraries were generated for each sampling point. For sampling points WAF_{t=7d}, WAF_{t=42d}, CEWAF_{t=7d}, and CEWAF_{t=42d} we generated duplicate transcriptomic libraries. The present study revealed the mechanistic response of the microbial community through assessment of DE genes mapped onto pangenomes. We identified the best reference genome for *Colwellia* and *Marinobacter*, as described above, and mapped metatranscriptomic libraries to all available complete and near to complete genomes from NCBI for these microorganisms (Supplementary Tables 1 and 2, Supplementary Data 1). The oil-only treatment showed the largest number of DE genes for *Marinobacter*, and the smallest number of DE genes for *Colwellia* (red and blue histograms in Figs. 1 and 2); all DE contrasts were calculated relative to the biotic control. In contrast, for *Colwellia*, the largest number of DE genes was observed in dispersant-amended treatments, while the smallest number of DE genes for *Marinobacter* occurred in dispersant-amended treatments. The distribution of upregulated genes ranged from 151 to 257 for *Marinobacter* and from 644 to 1472 for *Colwellia*.

In contrast, the distribution of downregulated genes ranged from 18 to 277 for *Marinobacter* and from 0 to 7 for *Colwellia*. Though both *Colwellia* and *Marinobacter* were present at low abundance initially (*Colwellia* at day 0: 16S rRNA gene abundance = 0.56%, metatranscriptomic abundance = 6.11%; *Marinobacter* at day 0: 16S rRNA gene abundance = 1.37%, metatranscriptomic abundance = 8.00%), and *Colwellia* was less abundant than *Marinobacter*; *Colwellia* exhibited a much stronger up regulation response to dispersants and/or hydrocarbons, while *Marinobacter* exhibited a comparable response regarding up and down regulation in response to dispersants and/or hydrocarbons.

To identify species-level functional traits that exhibited a transcriptional response to the imposed treatment regimes, we co-investigated the transcript abundance, activity, and the gene pool of *Colwellia* and *Marinobacter* populations across the microcosms. To this end, we extended the workflow implemented in *Anvi'o* for metapangenomics [49, 50] with metatranscriptomics, which we refer to as 'CA-Metatranscriptome', and we grouped gene clusters that contained at least one DE gene for downstream analyses (Figs. 1 and 2). Additionally, gene clusters were grouped into core and accessory gene clusters, based on their occurrence in all genomes or only a sub-set of genomes, respectively. In general, the core pangenome generates the broad ecological and phenotypic properties of a species [50–53]. In contrast, the accessory (or auxiliary) pangenome is associated with strain-specific, isolate-specific, rare- or common-to-subset fractions of a pangenome [56–59]. Whether the contribution of the accessory pangenome at the expression level is advantageous to phenotypes under selective conditions, especially in a harsh environment posed by oil and dispersants, is still an open question [60]. However, the accessory genome is known to promote adaptability in pathogens and may serve a similar function in other organisms [61]. By merging pangenomics, and metatranscriptomics in the same framework, CA-Metatranscriptomes enabled exploration of microbial in situ niche-contribution to relative transcriptional activity to identify the metabolic determinants of a particular microorganism's response to environmental perturbation.

The *Colwellia* CA-Metatranscriptome (Fig. 1) included 89,678 gene calls across 1,598 gene clusters. 40.9 % ($n = 653$) of the gene clusters were grouped into the core pangenome, while the rest ($n = 945$ or 59.1%) were present in the accessory pangenome. Most of the gene clusters contained DE genes with functional

annotation from KEGG/Kofam ($n = 69\%$) and from COGs ($n = 89.1\%$). The *Marinobacter* CA-Metatranscriptome (Fig. 2) included 98,405 gene calls across 826 gene clusters. For *Marinobacter*, 60 (7.3%) and 766 (92.7%) gene clusters were associated with the core and accessory CA-Metatranscriptome, respectively. Most of the gene clusters contained DE genes with functional annotation from KEGG/Kofam ($n = 59.7\%$) and from COGs ($n = 86.2\%$). So, while *Colwellia's* transcriptomic response was proportionally split between core (~41%, defined as DE clusters in *Colwellia* core pangenome $\times 100$ / total DE clusters in *Colwellia* pangenome) and accessory (~59%) gene clusters, *Marinobacter's* transcriptomic response was disproportionately dominated by the response of accessory (~93%) gene clusters. Previously, hydrocarbon degrading genes were reported in the accessory genome [62]. Furthermore, core and accessory responses in *Colwellia* did not impair housekeeping functions but stimulated several niche-specific accessory functions (see details below).

Colwellia exhibited a robust transcriptomic response in the core CA-Metatranscriptome (Fig. 1, Supplementary Data 2, Supplementary Results). Cell membrane biogenesis and inorganic ion transport categories were upregulated in the CEWAF and CEWAFN treatments. Additional genes related to fatty acid biosynthesis were upregulated in the CEWAFN treatment. Lipid metabolism and coenzyme metabolism were also upregulated in the CEWAF treatment. About 21% of upregulated genes in the core CA-Metatranscriptome of *Colwellia* were assigned to the top three COG categories J (translation), E (amino acid transport), and N (cell motility). In contrast, 23% of upregulated genes in the accessory CA-Metatranscriptome of *Colwellia* were in the top three COG categories T (signaling mechanisms), J and E. A sub-set of functions in the core CA-Metatranscriptome of *Colwellia* were upregulated in the dispersants-only treatment, reflecting a response to Corexit components rather than oil components (Fig. 1, Supplementary Data 2, Materials and Methods). The upregulated genes were involved in oxidative phosphorylation, energy and carbohydrate metabolism, two-component sensing systems, and recombination, suggesting metabolism of dispersant components to generate cellular energy. Furthermore, a collection of genes associated to carboxylic ester hydrolysis, sulfatase and C₁-C₂ bond dicarboxylic activity were upregulated with the dispersed treatments suggesting the metabolic potential for DOSS (i.e., dioctyl sodium sulfosuccinate) degradation in *Colwellia*, as suggested in the bulk microcosms by Seidel et al. [46].

Addition of CEWAFN stimulated specific biosynthetic pathways in *Colwellia's* accessory CA-Metatranscriptome (Fig. 1, Supplementary Data 2, Supplementary Results). Upregulated genes in this treatment included those involved in the synthesis of polyhydroxybutyrate, the most common polyhydroxyalkanoate, the largest known group of natural polyesters, and genes required for fatty acid degradation, such as *alkP* phosphoglycerate mutase, *ssuD* alkanesulfonate monooxygenase [63]. Genes involved in the degradation of aromatic hydrocarbons such as phenol 2-monooxygenase and catechol 2,3-dioxygenase genes, as well as membrane transport (TonB-dependent receptor) genes, were upregulated in all dispersant-containing treatments. For instance, the *fhuA* (formerly *tonA*), *fhuE*, and *fepA* genes were all upregulated in the DISP and CEWAFN treatments. These genes encode for iron-binding proteins involved in iron acquisition with targeted sensitivity towards ferric coprogen, ferric-rhodotorulic acid, ferrienterochelins and colicins [30, 64–66]. Interestingly, previous studies have shown the strong relationship between siderophore production and hydrocarbon (especially PAHs) degradation. Hydrocarbon degradation involves cleaving dioxygenases that contain iron in their active sites [67]. A marine *Vibrio* spp. isolated from the Deepwater Horizon oil spill produced ochrobactins (i.e., amphiphilic siderophores) with positive enhancement of hydrocarbon degradation [68]. Furthermore,

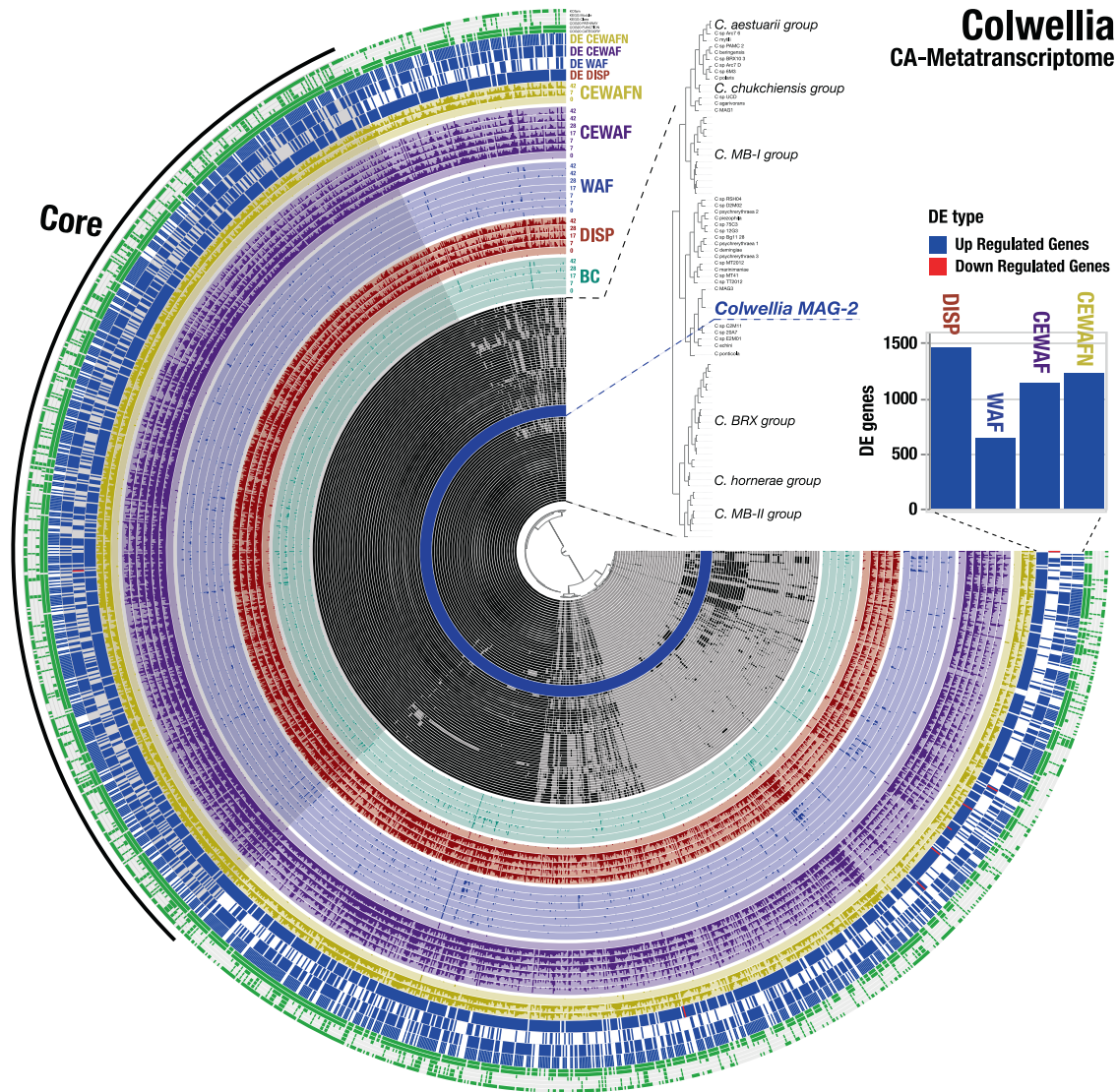


Fig. 1 Gene detection of metatranscriptomic reads in the context of the core and accessory genome in *Colwellia*. The 77 inner layers show the presence-absence of 1598 gene clusters with 89,678 genes that were identified in 75 *Colwellia* genomes and three *Colwellia* metagenomic assembled genomes (MAGs). An expanded dendrogram of the reference genomes based on the distribution of gene clusters using Euclidian distance and Ward's clustering is shown in the top-right. Some genome clusters were highlighted in gray blocks, and they are described in Supplementary Table 1. *Colwellia* MAG-2 (highlighted in blue) was the reference genome that recruited the largest fraction of transcriptomic reads among *Colwellia* genomes. Gene detection profiles of metatranscriptomic reads recruited by *Colwellia* MAG-2 are sorted, and color coded by the corresponding experimental treatments: Biotic control (BC), Dispersants (DISP), Water Accommodated oil Fraction (WAF), Chemically Enhanced Water Accommodated oil Fraction (CEWAF) and Chemically Enhanced Water Accommodated oil Fraction plus Nutrients (CEWAFN). Time is expressed in days. The next four layers show the presence-absence of DE genes with respect to the biotic control treatment that are shown in blue (up-regulation) and red (downregulation). A stacked-bar diagram on the right describes the DE gene counts across treatments following the color code for each of the treatments. The next six layers describe the gene clusters annotations in terms of Kofam, KEGG module, KEGG class, COG20 pathway, COG20 function and COG20 category. A detailed description of these DE genes is shown in Supplementary Data 2.

recent studies have shown clear correlation between siderophores and hydrocarbon degradation involving glutathione mediated stress responses (especially targeting reactive oxygen species), iron-driven co-metabolism, and surface area membrane activity [69–71]. Taken together, these results illustrate a sophisticated, niche specific response for *Colwellia* at the genus level.

Marinobacter was more transcriptionally active in the first week of the experiment (Fig. 2, Supplementary Data 2, Supplementary Results, Peña-Montenegro et al. submitted [47]), consistent with patterns of oil biodegradation when substrate was not limiting [18, 46]. The core of the *Marinobacter* CA-Metatranscriptome exhibited upregulation in response to the oil-only treatment, where 30% of these genes subscribed to repair protein genes

(COG category L), chaperon protein genes (COG category O) and signal transduction associated genes (COG category T). Additionally, we observed increased transcription of a group of house-keeping genes in the COG categories H (coenzyme metabolism), and E (amino acid transport). The reduced contribution of the core CA-Metatranscriptome relative to the accessory CA-Metatranscriptome to oil-only treatments, and the even further limited response to dispersant-amended treatments suggest highly specific adaptations of *Marinobacter* sp. C18 to oil exposure and suggest that these traits are not shared with other *Marinobacter* species.

In the accessory CA-Metatranscriptome of *Marinobacter* in the WAF treatment (Fig. 2, Supplementary Data 2, Supplementary

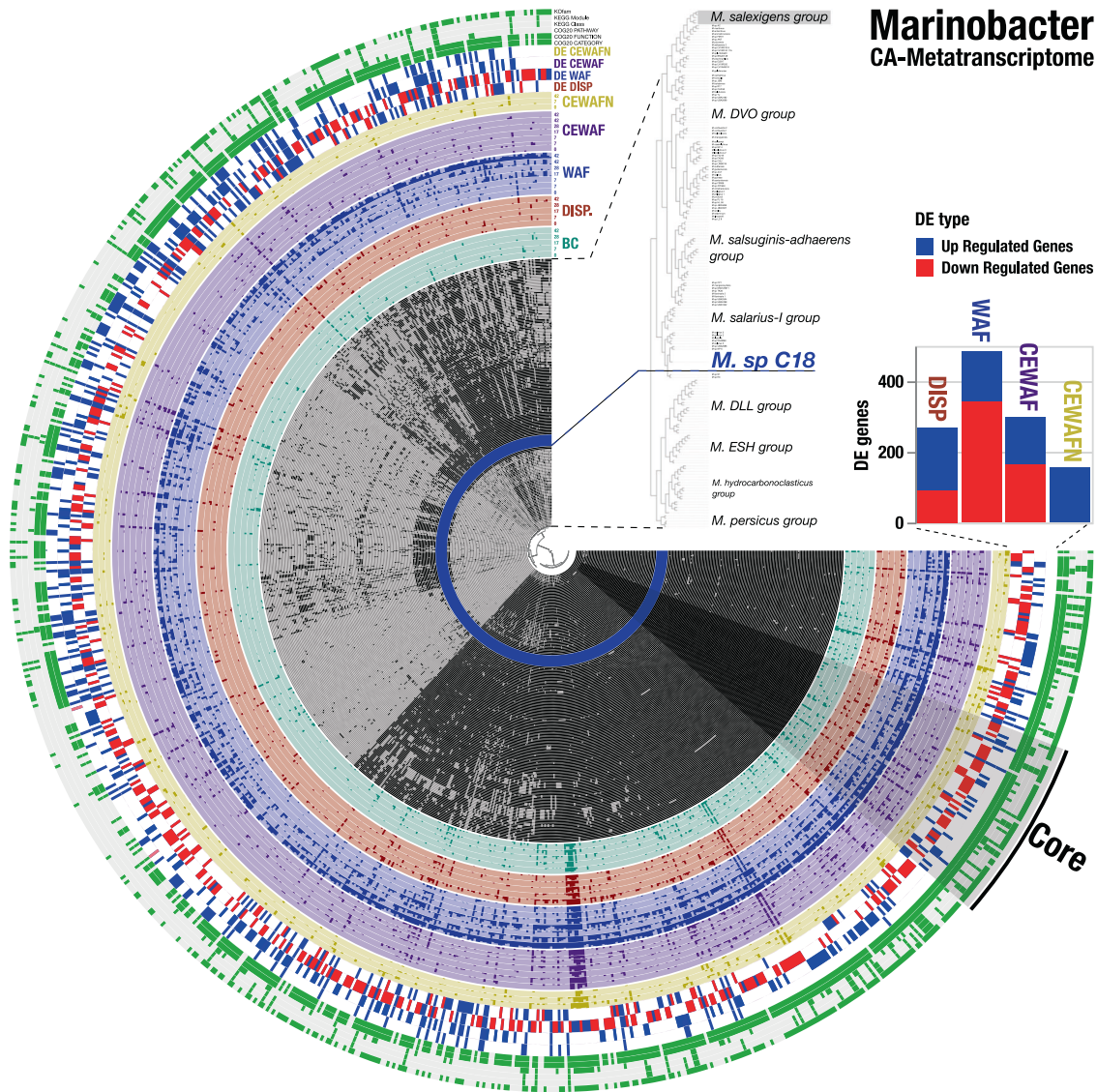


Fig. 2 Gene detection of metatranscriptomic reads in the context of the core and accessory genome in *Marinobacter*. The 171 inner layers show the presence-absence of 826 gene clusters with 98,405 genes that were identified in 180 *Marinobacter* genomes. An expanded dendrogram of the reference genomes based on the distribution of gene clusters using Euclidian distance and Ward's clustering is shown in the top-right. Some genome clusters were highlighted in gray blocks, and they are described in Supplementary Table 2. We defined the following groups based on their relevant representative species: **DVO** including *M. daqiaoensis*, *M. vulgaris*, and *M. orientalis*; **DLL** including *M. daepoensis*, *M. litoralis*, and *M. lutaensis*; and **ESH** including *M. excellens*, *M. shengliensis*, and *M. halophilus*. *Marinobacter* sp. C18 (highlighted in blue) was the reference genome that recruited the largest fraction of transcriptomic reads among *Marinobacter* genomes. Gene detection profiles of metatranscriptomic reads recruited by *Marinobacter* sp. C18 are sorted, and color coded by the corresponding experimental treatment: Biotic control (BC), Dispersants (DISP), Water-accommodated oil fraction (WAF), chemically enhanced water accommodated oil fraction (CEWAF) and chemically enhanced water accommodated oil fraction plus nutrients (CEWAFN). Time is expressed in days. The next four layers show the presence-absence of DE genes with respect to the biotic control treatment are shown in blue (up-regulation) and red (down-regulation). A stacked-bar diagram on the right describes the DE genes counts across treatments following the color code for each of the treatments. The next six layers describe the gene clusters annotations in terms of Kofam, KEGG module, KEGG class, COG20 pathway, COG20 function and COG20 category. A detailed description of these DE genes is shown in Supplementary Data 2.

Results), upregulated genes included essential genes involved in carbon and lipid metabolism, stress associated chaperone genes, fatty acid, and aromatic carbohydrate degradation genes. The latter genes may reflect metabolism of aromatic and xenobiotic intermediates found in the oil-derived carbon pool. Genes involved in chemotaxis, sensor kinases, ion homeostasis, and type IV pilus assembly were additionally upregulated. These physiological characteristics enable *Marinobacter* to respond efficiently to oil exposure, allowing these microorganisms to sense and move towards available substrate. *Marinobacter aquaeoli*, *Pseudomonas fluorescens*, and *Shewanella oneidensis*, which are rapid

responders to environmental perturbations, showed similar genomic features as the *Marinobacter* sp. C18 accessory CA-Metatranscriptome. Interestingly, these genomic features were mapped to transposable genomic sequences, which are typically associated to the accessory pangenome [42].

To compare the architecture of metabolic reaction pathways that were upregulated across the treatments, we mapped KEGG orthology numbers into the iPath 3 module [72] (Fig. 3). Fatty acid biosynthesis routes from acetate to medium-chain fatty acyl-CoA molecules were upregulated for *Marinobacter* (in the WAF treatment) and *Colwellia* (in the dispersants treatment).

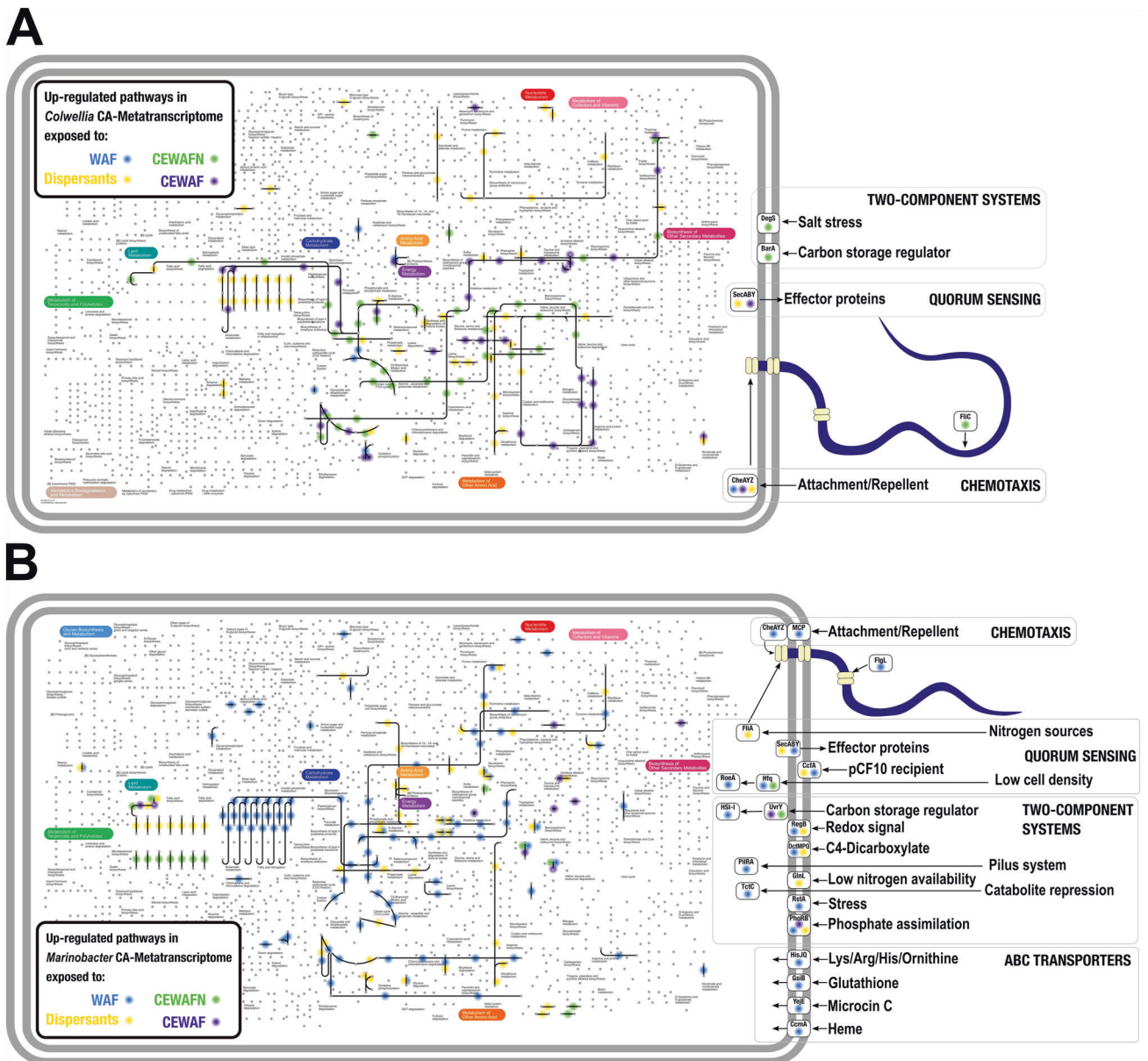


Fig. 3 Up-regulation maps of *Colwellia* MAG-2 and *Marinobacter* sp. C18 in simulated *Deepwater Horizon* microcosms. Bold lines indicate up-regulated reactions or components in **A** *Colwellia* MAG-2 and **B** *Marinobacter* sp. C18, under the exposure of WAF (blue), CEWAF (purple), CEWAFN (green), and Dispersants (yellow). Pathway visualization was generated using the iPath 3 module based on KEGG orthology numbers. An interactive version of the metabolic map is available at <https://pathways.embl.de/ipath3.cgi?s=xqYg5kHHCsxGqWy8yJ>.

Additionally, routes involving the TCA cycle (i.e., citrate – isocitrate – *cis*-aconitate conversion EC 4.2.1.3), and concomitant glycine biosynthesis of via alanine–glyoxylate transamination (EC 2.6.1.44), occurred in *Marinobacter* in the WAF treatment, and *Colwellia* in the CEWAFN treatment. Additional overlapping upregulated reactions between *Colwellia* and *Marinobacter* occurred in the oxidative phosphorylation routes and in purine metabolism. These observations matched unique functional trends over time for the *Colwellia* and *Marinobacter* communities suggesting different functional interactions in the WAF versus dispersant-amended treatments.

DISCUSSION

Colwellia signatures in the CA-Metatrascriptomes indicated opportunistic behavior in response to the chemical exposure regime and exposure time. Opportunitrophs typically show the

plasticity to utilize a broad spectrum of substrates, including harmful compounds found in crude oil, without becoming exclusively dependent on one source [73–75]. In addition, the opportunistic behavior is not only defined by their energy and carbon source, but also by ability to thrive in the face of variable spatial, temporal and ecological stressors [42, 76, 77]. *Colwellia* MAG-2 DE genes varied across the dispersants-only, CEWAF and CEWAFN treatments in distinct patterns with a smaller contribution from the accessory CA-Metatrascriptome in comparison to *Marinobacter*. Therefore, *Colwellia* showed a broader niche breath, resembling a more opportunistic lifestyle.

Although the upregulation of cytochrome P450 alkane hydroxylase genes in the *Colwellia* MAG-2 CA-Metatrascriptome was not observed in any treatment, we did observe a complex network of signaling, biogenesis, and regulation of cytochrome genes responding to oil-only (such as *cyoB*, and *ccmF*) versus dispersed treatments (such as *cyoC*, *ccmABE1*, *ccaA*, and *ccsA*), suggesting

WAF-dependent activation of carbohydrate oxidation. For instance, the role of the cytochrome ubiquinol oxidase (Cyo) in the modulation of hydrocarbon degradation via activation of cytochrome P450 hydroxylase genes has been documented [78]. Furthermore, multiple alkane utilization pathways can coexist, even involving unknown putative monooxygenases. Currently, the coordination mechanism is unclear [79]. Thus, the identification of predicted cytochrome P450 alkane hydroxylase genes is needed to better understand the mechanism of alkane hydroxylases in *Colwellia*.

Nutrient-dependent expression shifts in *Colwellia* MAG-2 were observed in genes involved in lipid metabolism (i.e., *fadEM*, *phaC*), hydrocarbon degradation (i.e., *xylE*, *alkP*) and TonB-dependent receptor genes. Our observations suggest that TonB-dependent receptors could be involved in the uptake of nutrients and downstream nutrient-associated benefits to adapt to oil/dispersant exposure. Previous studies have concluded that expression of TonB dependent receptor genes is key for opportunistic adaptations to harsh environments [30, 64]. Most importantly, a large proportion of the *Colwellia* MAG-2 transcriptomic responses were associated with the core CA-Metatranscriptome, reflecting an opportunistic response at genus level (Fig. 1). These results show that *Colwellia*, at the genus level, do not dominate a particular niche. Instead, they employ genomic plasticity as a strategy to survive under conditions of Corexit exposure.

The response of *Marinobacter* illustrated the profile of a specialized and well-adapted oil degrader. In this context, a specialized oil degrader resembles the behavior of r strategists and copiotrophs. These organisms would be able to grow fast and adapt to transient, unstable and uncrowded patches of rich resources, until they die or become dormant due to substrate exhaustion [42]. Exposure to WAF led *Marinobacter* sp. C18 to up-regulate the transcription of β -oxidation genes (*fadAH*, *dctMP*), likely indicating increased hydrocarbon degradation activity [80]. Additionally, a wide variety of DE genes in *Marinobacter* sp. C18 were involved in interactions with its environment, such as chemotaxis genes (*cheCY*, *mcp*, *motAB*), flagellar genes (*flgLJ*, *flaA1*, *fliD*), and genes for type IV pilus assembly (*pilBCF*, *tadD*). These observations are consistent with the description of extracellular sensing and/or aggregate formation linked to hydrocarbon degradation [38, 39, 81]. The majority of WAF-specific responses were adaptations at the species level, i.e., the response was generated by the accessory CA-Metatranscriptome of *Marinobacter* sp. C18, rather than at the genus level, i.e., generated by the core CA-Metatranscriptome of *Marinobacter* sp. C18 (Fig. 2). These results match the lifestyle of a microorganism seeking an adaptive advantage in a changing environment; which usually correlates with a small core and an open pangenome [60]. *Marinobacter* is well-adapted to respond swiftly and efficiently to dissolved oil exposures in its environment.

Nutrient availability may limit a microorganism's ability to respond to substrate enrichment or allow a microorganism to accelerate its metabolic response and/or tolerate otherwise stressful conditions. The application of oil, dispersants and nutrients – i.e., the CEWAFN treatment – shifted the upregulation map for *Colwellia* MAG-2 and *Marinobacter* sp. C18 relative to other treatments and over time (Fig. 3). In the WAF treatment, a diverse and complex response in *Marinobacter* sp. C18 was evident by DE genes involving chemotaxis, membrane two-component sensors, ABC transporters, secretion systems, quorum sensing components, fatty acid and aromatic hydrocarbon degradation routes, butyrate, propionate and glutamate assimilation, downstream biosynthesis and metabolism of nucleotides, as well as cofactors and vitamins utilized for biosynthesis. These upregulated metabolisms reflected an inherent ability to sense and metabolize hydrocarbons.

In contrast, *Colwellia* MAG-2 exhibited striking changes in the reaction map, including utilization of glutamine for the

biosynthesis of inosine, a precursor for adenosine and nucleotides, consumption of L-glutamate for the biosynthesis of heme cofactors, possibly involved in the biosynthesis of P450 cytochromes and nitrogen metabolism via transcription of [Fe-S] cluster assembly proteins in the WAF treatment. Sequence alignments in PAH degrading *Mycobacterium* isolates displayed the following operon structure: a cytochrome P450 monooxygenase (CYP151 – *pipA*, CYP150 or CYP51), a ferredoxin (*fdx*), and a glutamine synthetase (*glnA*) [82]. The possibility of a dispersant-inducible operon with a similar arrangement in *Colwellia* is hypothesized based on the significant co-expression of these genes.

Colwellia's response resembled a modularized network tuning reactions to specific changes in the environment (Fig. 3). Most of the altered reaction modules were not associated with membrane or signaling systems that would respond to external stimuli. Responses at the expression level involving internal metabolic modules in *Colwellia* have been documented [65, 66] and support our hypothesis that *Colwellia* acts as an individualistic opportunist following exposure of oil and dispersants. In contrast, the complex interaction with the surrounding environment observed in *Marinobacter* suggests community-driven, environment-dependent metabolic alterations combined with a rapid and effective ability to respond to hydrocarbon exposure at a species level.

Previous studies either provide support against [37, 40] or, interestingly, in favor of [41, 42, 75] the perception of *Marinobacter* as an opportunist. Our transcriptomic datasets and CA-Metatranscriptomic analysis show different expression responses across the metabolic map and suggest very different responses in a specialist-vs-opportunistic framework between *Marinobacter* and *Colwellia* (Fig. 3). In the CEWAFN treatment, *Colwellia* exhibited upregulation for the biosynthesis of glycerophospholipids, and the metabolism of amino acids, suggesting cell membrane and chaperone maintenance to combat chemical stress. Oil and dispersants are known to induce cytotoxic effects in the bacterial membrane [83], including changes in (glycerophospho)lipids [84]. For instance, the impairment of glycerophospholipids in the bacterial membrane was reported to be linked to global inhibition of dehydrogenase activity, causing reduced ATP levels and increased reactive oxidative species stress. Downstream consequences included increased amino acid utilization, possibly compensating for increased environmental susceptibility and the reduced ability to trigger membrane adaptation and maintenance [85].

Although the response of *Marinobacter* and *Colwellia* were similar in the CEWAFN treatment with regard to fatty acid biosynthesis, *Marinobacter* sp. C18 showed specific upregulation of the carbon-storage regulator CsrA, cell density sensors, and molybdenum cofactor biosynthesis. This similar response indicates that nutrients improved the fitness of *Marinobacter* in this treatment. CsrA has been described as a conserved global regulator to control central carbon pathways, motility, biofilm formation, and pathogenicity [86]. *Marinobacter*'s response to nutrients aligned with previous studies [87] where the upregulation of CsrA matched the upregulation of fatty acid biosynthesis pathway, nutrient uptake systems, maintenance of the cell envelope integrity, and regulation of iron uptake. Interestingly, *Colwellia* MAG-2 showed CsrA upregulation in treatments without nutrients, highlighting its ability to fine-tune catabolic and anabolic pathways so as to optimize energy use in response to changing conditions.

The CA-Metatranscriptome approach provided valuable insight to the complex interactions between members of the microbial community and their surrounding environment that occur in response to perturbation and confirms the hypothesis of Goyal (2018) that accessory genome components promote response and adaptability in microbial populations [88]. A complex suite of responses in *Colwellia* and *Marinobacter* CA-Metatranscriptomes

revealed unique aspects of their metabolic capabilities that explained treatment-specific transcript abundances and activities. *Colwellia* exhibited an opportunistic response to organic carbon enrichment, while *Marinobacter* displayed a fine-tuned response to oil pollution. Genomic and transcriptomic plasticity promoted the success of one versus the other across the treatment regime and underscores the role of specialist-vs-opportunistic components of marine microbiomes, revealing how and why specific organisms respond to particular conditions through employment of metabolic cassettes associated with core versus accessory pangenomes.

The assessment of species-level gene content variation (i.e., the accessory pangenome) in microbial populations is needed to better understand the ecological and metabolic plasticity and how accessory genes facilitate adaptation to environmental perturbations. Previous studies have documented the core and accessory pangenome [56–59], but it is unclear whether results from the comparison of species-level pangenomes acquired from different habitats and samples reflect the ecological fitness of natural populations to respond to environmental perturbations. The present results provide further evidence that accessory genes are associated to niche-specific adaptations (i.e., oil-only responses in the *Marinobacter* accessory CA-Metatranscriptome) while core genes aligned with essential functions (i.e., a broad range response in the *Colwellia* core CA-Metatranscriptome). Interestingly, core genes can have higher recombination rates than accessory genes, as recently reported [89], which may suggest a more effective selection in *Colwellia* by environmental perturbations.

Most of the observed responses to WAF in *Marinobacter* arose from species level responses, illustrating that distinct ecotypes thrived under oil-infused conditions. In contrast, for *Colwellia*, both core and accessory CA-Metatranscriptome showed transcriptomic signals, mostly in response to dispersant addition. *Colwellia* and *Marinobacter* in dispersant-amended and WAF treatments, respectively, exhibited functional responses that followed different treatment-dependent trajectories along the metabolic map (Fig. 3). The CA-Metatranscriptome data underscores the fact that the decrease of *Marinobacter* in dispersant treatments stems from metabolic impairment – namely that dispersants inhibited *Marinobacter's* metabolism, which decreased its abundance. In contrast, dispersants stimulated *Colwellia's* metabolism and promoted its growth and biomass accumulation. Competitive interactions between *Marinobacter* and *Colwellia* may have further contributed to the observed selection against *Marinobacter* in dispersant treatments, but the previously reported patterns derive primarily from physiological effects that became clear in the metatranscriptomic approach of the current study.

Recent advances underscore the importance of species-specific functional responses in modulating ecological and functional signatures on the ecosystem level. Indeed, the response and subsequent recovery of microbial communities to perturbation is a key determinant of ecosystem function across a broad range of disturbances [90, 91]. Unraveling the ways through which species-specific responses to perturbation are manifest in functional change is challenging [92], though possible if focusing on a single disturbance [93]. The approach applied to the data herein reveal that unique species-specific drivers enable community-level responses to disturbance, and the CA-Metatranscriptome approach provides a blueprint for assessing microbial responses to environmental perturbation across Earth's ecosystems, paving the way to assess how and why key microbial players respond to disturbance in a particular fashion.

DATA AVAILABILITY

All scripts are found on Github (<https://github.com/biotemon/K2015>). Anvi'o procedures for *Colwellia* and *Marinobacter* CA-Metatranscriptome curation are

explicit in Supplementary Data 3 and 4, respectively. Raw sequencing reads generated for this study can be found in the Sequence Read Archive under the BioProject [PRJNA640753](https://www.ncbi.nlm.nih.gov/bioproject/PRJNA640753). We made available Anvi'o CA-Metatranscriptome files, and Anvi'o summarized profiles in the Open Science Framework repository at <https://doi.org/10.17605/OSF.IO/FU9BW>.

REFERENCES

- Worden AZ, Follows MJ, Giovannoni SJ, Wilken S, Zimmerman AE, Keeling PJ. Rethinking the marine carbon cycle: Factoring in the multifarious lifestyles of microbes. *Science*. 2015;347:1257594–1257594.
- Moran MA. The global ocean microbiome. *Science*. 2015;350:aac8455–aac8455.
- Sunagawa S, Coelho LP, Chaffron S, Kultima JR, Labadie K, Salazar G, et al. Structure and function of the global ocean microbiome. *Science*. 2015;348:1261359–1261359.
- Louca S, Palfrey LW, Doebeli M. Decoupling function and taxonomy in the global ocean microbiome. *Science*. 2016;353:1272–7.
- Sogin ML, Morrison HG, Huber JA, Welch DM, Huse SM, Neal PR, et al. Microbial diversity in the deep sea and the underexplored 'rare biosphere'. *Proc Natl Acad Sci USA*. 2006;103:12115–20.
- Lynch DJ, Neufeld JD. Ecology and exploration of the rare biosphere. *Nat Rev Microbiol*. 2015;13:217–29.
- Kleindienst S, Grim S, Sogin M, Bracco A, Crespo-Medina M, Joye SB. Diverse, rare microbial taxa responded to the Deepwater Horizon deep-sea hydrocarbon plume. *ISME J*. 2016;10:400–15.
- Blount K, Jones C, Walsh D, Gonzalez C, Shannon WD. Development and validation of a novel microbiome-based biomarker of post-antibiotic dysbiosis and subsequent restoration. *Front Microbiol*. 2022;12:781275.
- Bardgett RD, Caruso T. Soil microbial community responses to climate extremes: resistance, resilience and transitions to alternative states. *Phil Trans R Soc B*. 2020;375:20190112.
- Dove NC, Klingeman DM, Carrell AA, Cregger MA, Schadt CW. Fire alters plant microbiome assembly patterns: integrating the plant and soil microbial response to disturbance. *New Phytol*. 2021;230:2433–46.
- Zhang Q, Zhang Z, Lu T, Yu Y, Penuelas J, Zhu Y-G, et al. Gammaproteobacteria, a core taxon in the guts of soil fauna, are potential responders to environmental concentrations of soil pollutants. *Microbiome*. 2021;9:196.
- Biessy L, Pearman JK, Waters S, Vandergoes MJ, Wood SA. Metagenomic insights to the functional potential of sediment microbial communities in freshwater lakes. *MBMG*. 2022;6:e79265.
- De Anda V, Zapata-Peñasco I, Blaz J, Poot-Hernández AC, Contreras-Moreira B, González-Laffitte M, et al. Understanding the mechanisms behind the response to environmental perturbation in microbial mats: a metagenomic-network based approach. *Front Microbiol*. 2018;9:2606.
- Galand PE, Lucas S, Fagervold SK, Peru E, Pruski AM, Vétion G, et al. Disturbance increases microbial community diversity and production in marine sediments. *Front Microbiol*. 2016;7:1950.
- Cárdenas A, Raina J-B, Pogoreutz C, Räderer N, Bougoure J, Guagliardo P, et al. Greater functional diversity and redundancy of coral endolithic microbiomes align with lower coral bleaching susceptibility. *ISME J*. 2022;16:2406–20.
- Dang H, Klotz MG, Lovell CR, Sievert SM. Editorial: the responses of marine microorganisms, communities and ecofunctions to environmental gradients. *Front Microbiol*. 2019;10:115.
- Joye S, Kostka JE. Microbial genomics of the global ocean system: report on an American Academy of Microbiology (Academy), The American Geophysical Union (AGU), and The Gulf of Mexico Research Initiative (GoMRI) Colloquium held on 9 and 10 April 2019. Washington, DC: American Society for Microbiology; 2020.
- Kleindienst S, Seidel M, Ziervogel K, Grim S, Loftis K, Harrison S, et al. Chemical dispersants can suppress the activity of natural oil-degrading microorganisms. *Proc Natl Acad Sci USA*. 2015;112:14900–5.
- Crone TJ, Tolstoy M. Magnitude of the 2010 Gulf of Mexico oil leak. *Science*. 2010;330:634–634.
- National Commission on the BP Deepwater Horizon Oil Spill and Offshore Drilling. The use of surface and subsea dispersants during the BP Deepwater Horizon oil spill. 2011. Washington, D.C.; National Commission on the BP Deepwater Horizon Oil Spill and Offshore Drilling.
- Techtman SM, Zhuang M, Campo P, Holder E, Elk M, Hazen TC, et al. Corexit 9500 enhances oil biodegradation and changes active bacterial community structure of oil-enriched microcosms. *Appl Environ Microbiol*. 2017;83:e03462–16. e03462-16
- Hamdan L, Fulmer P. Effects of COREXIT® EC9500A on bacteria from a beach oiled by the Deepwater Horizon spill. *Aquat Microb Ecol*. 2011;63:101–9.
- Mason OU, Hazen TC, Borglin S, Chain PSG, Dubinsky EA, Fortney JL, et al. Metagenome, metatranscriptome and single-cell sequencing reveal microbial response to Deepwater Horizon oil spill. *ISME J*. 2012;6:1715–27.

24. Rodríguez-R LM, Overholt WA, Hagan C, Huettel B, Kostka JE, Konstantinidis KT. Microbial community successional patterns in beach sands impacted by the Deepwater Horizon oil spill. *ISME J*. 2015;9:1928–40.
25. Dubinsky EA, Conrad ME, Chakraborty R, Bill M, Borglin SE, Hollibaugh JT, et al. Succession of hydrocarbon-degrading bacteria in the aftermath of the Deepwater Horizon oil spill in the Gulf of Mexico. *Environ Sci Technol*. 2013;47:10860–7.
26. Hazen TC, Dubinsky EA, DeSantis TZ, Andersen GL, Piceno YM, Singh N, et al. Deep-sea oil plume enriches indigenous oil-degrading Bacteria. *Science*. 2010;330:204–8.
27. Valentine DL, Kessler JD, Redmond MC, Mendes SD, Heintz MB, Farwell C, et al. Propane respiration jump-starts microbial response to a deep oil spill. *Science*. 2010;330:208–11.
28. Redmond MC, Valentine DL. Natural gas and temperature structured a microbial community response to the Deepwater Horizon oil spill. *Proc Natl Acad Sci USA*. 2012;109:20292–7.
29. Deming JW, Somers LK, Straube WL, Swartz DG, Macdonell MT. Isolation of an obligately barophilic bacterium and description of a new genus, *Colwellia* gen. nov. *Syst Appl Microbiol*. 1988;10:152–60.
30. Mason OU, Han J, Woyke T, Jansson JK. Single-cell genomics reveals features of a *Colwellia* species that was dominant during the Deepwater Horizon oil spill. *Front Microbiol*. 2014;5:332.
31. Carillo S, Casillo A, Pieretti G, Parrilli E, Sannino F, Bayer-Giraldi M, et al. A unique capsular polysaccharide structure from the psychrophilic marine bacterium *Colwellia psychrelythraea* 34H that mimics antifreeze (Glyco)proteins. *J Am Chem Soc*. 2015;137:179–89.
32. Lu L, Rughöft S, Straub D, Joye SB, Kappler A, Kleindienst S. Rhamnolipid biosurfactants enhance microbial oil biodegradation in surface seawater from the North Sea. *ACS EST Water*. 2023;3:2255–66.
33. Campeão ME, Reis L, Leomil L, de Oliveira L, Otsuki K, Gardinali P, et al. The Deep-sea microbial community from the Amazonian Basin associated with oil degradation. *Front Microbiol*. 2017;8:1019.
34. Tripathi L, Irorere VU, Marchant R, Banat IM. Marine derived biosurfactants: a vast potential future resource. *Biotechnol Lett*. 2018;40:1441–57.
35. Gauthier MJ, Lafay B, Christen R, Fernandez L, Acquaviva M, Bonin P, et al. *Marinobacter hydrocarbonoclasticus* gen. nov., sp. nov., a new, extremely halotolerant, hydrocarbon-degrading marine bacterium. *Int J Syst Evol Microbiol*. 1992;42:568–76.
36. Ivanova EP, Ng HJ, Webb HK, Feng G, Oshima K, Hattori M, et al. Draft genome sequences of *Marinobacter similis* A3d10T and *Marinobacter salarius* R9SW1T. *Genome Announc*. 2014;2:e00442–14. 2/3/e00442-14
37. Rughöft S, Vogel AL, Joye SB, Gutierrez T, Kleindienst S. Starvation-dependent inhibition of the hydrocarbon degrader *Marinobacter* sp. TT1 by a chemical dispersant. *JMSE*. 2020;8:925.
38. Mounier J, Camus A, Mitteau I, Vaysse P-J, Goulas P, Grimaud R, et al. The marine bacterium *Marinobacter hydrocarbonoclasticus* SP17 degrades a wide range of lipids and hydrocarbons through the formation of oleolytic biofilms with distinct gene expression profiles. *FEMS Microbiol Ecol*. 2014;90:816–31.
39. Ennouri H, d'Abzac P, Hakil F, Branchu P, Naïtali M, Lomenech A, et al. The extracellular matrix of the oleolytic biofilms of *Marinobacter hydrocarbonoclasticus* comprises cytoplasmic proteins and T2SS effectors that promote growth on hydrocarbons and lipids. *Environ Microbiol*. 2017;19:159–73.
40. Eddie BJ, Malanoski AP, Onderko EL, Phillips DA, Glaven SM. *Marinobacter atlanticus* electrode biofilms differentially regulate gene expression depending on electrode potential and lifestyle. *Biofilm*. 2021;3:100051.
41. Pinto J, Lami R, Krasovec M, Grimaud R, Urios L, Lupette J, et al. Features of the opportunistic behaviour of the marine bacterium *Marinobacter algicola* in the microalga *Ostreococcus tauri* phycosphere. *Microorganisms*. 2021;9:1777.
42. Singer E, Webb EA, Nelson WC, Heidelberg JF, Ivanova N, Pati A, et al. Genomic potential of *Marinobacter aquaeolei*, a biogeochemical “opportunotroph”. *Appl Environ Microbiol*. 2011;77:2763–71.
43. Dombrowski N, Donahoe JA, Gutierrez T, Seitz KW, Teske AP, Baker BJ. Reconstructing metabolic pathways of hydrocarbon-degrading bacteria from the Deepwater Horizon oil spill. *Nat Microbiol*. 2016;1:16057.
44. Yuan J. The diversity of PAH-degrading bacteria in a deep-sea water column above the Southwest Indian Ridge. *Front Microbiol*. 2015;6:853.
45. Tremblay J, Yergeau E, Fortin N, Cobanli S, Elias M, King TL, et al. Chemical dispersants enhance the activity of oil- and gas condensate-degrading marine bacteria. *ISME J*. 2017;11:2793–808.
46. Seidel M, Kleindienst S, Dittmar T, Joye SB, Medeiros PM. Biodegradation of crude oil and dispersants in deep seawater from the Gulf of Mexico: Insights from ultra-high resolution mass spectrometry. *Deep Sea Res Part II*. 2016;129:108–18.
47. Peña-Montenegro TD, Kleindienst S, Allen AE, Eren AM, McCrow JP, Sánchez-Calderón JD, et al. Metatranscriptomic response of deep ocean microbial populations to infusions of oil and/or synthetic chemical dispersant. *OSF*. <https://doi.org/10.17605/OSF.IO/FU9BW>. Accessed 5 Sep 2023.
48. McCarren J, Becker JW, Repeta DJ, Shi Y, Young CR, Malmstrom RR, et al. Microbial community transcriptomes reveal microbes and metabolic pathways associated with dissolved organic matter turnover in the sea. *Proc Natl Acad Sci USA*. 2010;107:16420–7.
49. Delmont TO, Eren AM. Linking pangenomes and metagenomes: the *Prochlorococcus* metapangenome. *PeerJ*. 2018;6:e4320.
50. Eren AM, Kiefl E, Shaiber A, Veseli I, Miller SE, Schechter MS, et al. Community-led, integrated, reproducible multi-omics with anvio. *Nat Microbiol*. 2020;6:3–6.
51. Howe A, Stopnisek N, Dooley SK, Yang F, Grady KL, Shade A. Seasonal activities of the phyllosphere microbiome of perennial crops. *Nat Commun*. 2023;14:1039.
52. Dijkhuizen LW, Tabatabaei BES, Brouwer P, Rijken N, Buijs VA, Güngör E, et al. Far-red light-induced azolla filiculoides symbiosis sexual reproduction: responsive transcripts of symbiont nostoc azollae encode transporters whilst those of the fern relate to the angiosperm floral transition. *Front Plant Sci*. 2021;12:693039.
53. Alio I, Gudzuhan M, Pérez García P, Danso D, Schoelmerich MC, Mamat U, et al. Phenotypic and transcriptomic analyses of seven clinical *Stenotrophomonas maltophilia* isolates identify a small set of shared and commonly regulated genes involved in the biofilm lifestyle. *Appl Environ Microbiol*. 2020;86:e02038–20.
54. Wagner GP, Kin K, Lynch VJ. Measurement of mRNA abundance using RNA-seq data: RPKM measure is inconsistent among samples. *Theory Biosci*. 2012;131:281–5.
55. Love MI, Huber W, Anders S. Moderated estimation of fold change and dispersion for RNA-seq data with DESeq2. *Genome Biol*. 2014;15:550.
56. Conrad RE, Viver T, Gago JF, Hatt JK, Venter SN, Rossello-Mora R, et al. Toward quantifying the adaptive role of bacterial pangenomes during environmental perturbations. *ISME J*. 2022;16:1222–34.
57. Tettelin H, Riley D, Cattuto C, Medini D. Comparative genomics: the bacterial pangenome. *Curr Opin Microbiol*. 2008;11:472–7.
58. Medini D, Donati C, Tettelin H, Masignani V, Rappuoli R. The microbial pangenome. *Curr Opin Genet Dev*. 2005;15:589–94.
59. Vernikos G, Medini D, Riley DR, Tettelin H. Ten years of pan-genome analyses. *Curr Opin Microbiol*. 2015;23:148–54.
60. Golcz AA, Bayer PE, Bhalla PL, Batley J, Edwards D. Pangenomics comes of age: from bacteria to plant and animal applications. *Trends Genet*. 2020;36:132–45.
61. Jackson RW, Vinatzer B, Arnold DL, Dorus S, Murillo J. The influence of the accessory genome on bacterial pathogen evolution. *Mobile Genet Elements*. 2011;1:55–65.
62. Somee MR, Amoozegar MA, Dastgheib SMM, Shavandi M, Maman LG, Bertilsson S, et al. Genome-resolved analyses show an extensive diversification in key aerobic hydrocarbon-degrading enzymes across bacteria and archaea. *BMC Genomics*. 2022;23:690.
63. My L, Rekoske B, Lemke JJ, Viala JP, Gourse RL, Bouveret E. Transcription of the *Escherichia coli* fatty acid synthesis operon *fabH* is directly activated by FadR and Inhibited by ppGpp. *J Bacteriol*. 2013;195:3784–95.
64. Mudge MC, Nunn BL, Firth E, Ewert M, Hales K, Fondrie WE, et al. Subzero, saline incubations of *Colwellia psychrelythraea* reveal strategies and biomarkers for sustained life in extreme icy environments. *Environ Microbiol*. 2021;23:3840–66.
65. Czajka JJ, Abernathy MH, Benites VT, Baidoo EEK, Deming JW, Tang YJ. Model metabolic strategy for heterotrophic bacteria in the cold ocean based on *Colwellia psychrelythraea* 34H. *Proc Natl Acad Sci USA*. 2018;115:12507–12.
66. Christiansen L, Pathiraja D, Bech PK, Schultz-Johansen M, Hennessy R, Teze D, et al. A multifunctional polysaccharide utilization gene cluster in *Colwellia echini* encodes enzymes for the complete degradation of κ-carrageenan, ι-carrageenan, and hybrid β/κ-carrageenan. *mSphere*. 2020;5:e00792–19.
67. Harayama S, Rekik M. Bacterial aromatic ring-cleavage enzymes are classified into two different gene families. *J Biol Chem*. 1989;264:15328–33.
68. Gauglitz JM, Zhou H, Butler A. A suite of citrate-derived siderophores from a marine *Vibrio* species isolated following the Deepwater Horizon oil spill. *J Inorganic Biochem*. 2012;107:90–95.
69. Zhang N, Dan A, Chao Y, Li H-Y, Li C, Lin Q, et al. Mechanism of polycyclic aromatic hydrocarbons degradation in the rhizosphere of *Phragmites australis*: Organic acid co-metabolism, iron-driven, and microbial response. *Environ Pollut*. 2023;327:121608.
70. Liu J, Zhang A-N, Liu Y-J, Liu Z, Liu Y, Wu X-J. Analysis of the mechanism for enhanced pyrene biodegradation based on the interactions between iron-ions and *Rhodococcus ruber* strain L9. *Ecotoxicol Environ Saf*. 2021;225:112789.
71. Singha LP, Sinha N, Pandey P. Rhizoremediation prospects of Polyaromatic hydrocarbon degrading rhizobacteria, that facilitate glutathione and glutathione-S-transferase mediated stress response, and enhance growth of rice plants in pyrene contaminated soil. *Ecotoxicol Environ Saf*. 2018;164:579–88.
72. Darzi Y, Letunic I, Bork P, Yamada T. iPath3.0: interactive pathways explorer v3. *Nucleic Acids Res*. 2018;46:W510–W513.

73. Kostka JE, Prakash O, Overholt WA, Green SJ, Freyer G, Canion A, et al. Hydrocarbon-degrading bacteria and the bacterial community response in Gulf of Mexico beach sands impacted by the deepwater horizon oil spill. *Appl Environ Microbiol.* 2011;77:7962–74.
74. Handley KM, Lloyd JR. Biogeochemical implications of the ubiquitous colonization of marine habitats and redox gradients by *Marinobacter* species. *Front Microbiol.* 2013;4:136.
75. Sebastián M, Estrany M, Ruiz-González C, Forn I, Sala MM, Gasol JM, et al. High growth potential of long-term starved deep ocean opportunistic heterotrophic bacteria. *Front Microbiol.* 2019;10:760.
76. Moran MA, Buchan A, González JM, Heidelberg JF, Whitman WB, Kiene RP, et al. Genome sequence of *Silicibacter pomeroyi* reveals adaptations to the marine environment. *Nature.* 2004;432:910–3.
77. Polz MF, Hunt DE, Preheim SP, Weinreich DM. Patterns and mechanisms of genetic and phenotypic differentiation in marine microbes. *Phil Trans R Soc B.* 2006;361:2009–21.
78. Sevilla E, Yuste L, Moreno R, Rojo F. Differential expression of the three *Alcanivorax borkumensis* SK2 genes coding for the P450 cytochromes involved in the assimilation of hydrocarbons: Induction of *A. borkumensis* SK2 P450 cytochromes. *Environ Microbiol Rep.* 2017;9:797–808.
79. Wang W, Shao Z. The long-chain alkane metabolism network of *Alcanivorax dieselolei*. *Nat Commun.* 2014;5:5755.
80. Krivoruchko A, Zhang Y, Siewers V, Chen Y, Nielsen J. Microbial acetyl-CoA metabolism and metabolic engineering. *Metab Eng.* 2015;28:28–42.
81. Espinosa-Urgel M, Marqués S. New insights in the early extracellular events in hydrocarbon and lipid biodegradation. *Environ Microbiol.* 2017;19:15–18.
82. Brezna B, Kweon O, Stingley RL, Freeman JP, Khan AA, Polek B, et al. Molecular characterization of cytochrome P450 genes in the polycyclic aromatic hydrocarbon degrading *Mycobacterium vanbaalenii* PYR-1. *Appl Microbiol Biotechnol.* 2006;71:522–32.
83. Rughöft S, Jehmlich N, Gutierrez T, Kleindienst S. Comparative proteomics of *Marinobacter* sp. TT1 reveals Corexit impacts on hydrocarbon metabolism, chemotactic motility and biofilm formation. *Microorganisms.* 2020;9:3.
84. Sikkema J, De Bont JA, Poolman B. Mechanisms of membrane toxicity of hydrocarbons. *Microbiol Rev.* 1995;59:201–22.
85. Rowlett VW, Mallampalli VKPS, Karlstaedt A, Dowhan W, Taegtmeier H, Margolin W, et al. Impact of membrane phospholipid alterations in *Escherichia coli* on cellular function and bacterial stress adaptation. *J Bacteriol.* 2017;199:e00849–16.
86. Berndt V, Beckstette M, Volk M, Dersch P, Brönstrup M. Metabolome and transcriptome-wide effects of the carbon storage regulator A in enteropathogenic *Escherichia coli*. *Sci Rep.* 2019;9:138.
87. Potts AH, Vakulskas CA, Pannuri A, Yakhnin H, Babitzke P, Romeo T. Global role of the bacterial post-transcriptional regulator CsrA revealed by integrated transcriptomics. *Nat Commun.* 2017;8:1596.
88. Goyal A. Metabolic adaptations underlying genome flexibility in prokaryotes. *PLoS Genet.* 2018;14:e1007763.
89. Preska Steinberg A, Lin M, Kussell E. Core genes can have higher recombination rates than accessory genes within global microbial populations. *eLife.* 2022;11:e78533.
90. Allison SD, Martiny JBH. Resistance, resilience, and redundancy in microbial communities. *Proc Natl Acad Sci USA.* 2008;105:11512–9.
91. Philippot L, Griffiths BS, Langenheder S. Microbial community resilience across ecosystems and multiple disturbances. *Microbiol Mol Biol Rev.* 2021;85:e00026–20.
92. Rillig MC, Ryo M, Lehmann A, Aguilar-Trigueros CA, Buchert S, Wulf A, et al. The role of multiple global change factors in driving soil functions and microbial biodiversity. *Science.* 2019;366:886–90.
93. Shade A, Peter H, Allison SD, Baho DL, Berga M, Bürgmann H, et al. Fundamentals of microbial community resistance and resilience. *Front Microbiol.* 2012;3:417.

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AUTHOR CONTRIBUTIONS

Conceptualization: SBJ, SK, TDPM and JA. Methodology: SBJ, SK, TDPM, AME, AEA and JPM. Investigation: TDPM, SBJ, JA and AME. Visualization: TDPM and AME. Supervision: SBJ, JA, and JDSC. Writing—original draft: TDPM. Writing—review & editing: SBJ, TDPM, SK, and all co-authors.

COMPETING INTERESTS

The authors declare no competing interests.

ADDITIONAL INFORMATION

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